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Sustainable Reclaimed Water Management by Subsurface Drip Irrigation System: a study case for forage production

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Abstract In the Canary Islands the price fluctuation of groundwater (0.3 to $0.7 \notin m^3$), which is a private resource, often limits the recovery of the agricultural investments. Availability of alternative resources (reclaimed water: RW, desalinized seawater: DW and desalinized reclaimed water: DRW) at a foreseeable price will allow people to design the optimal infrastructures adapted to their needs The pumping cost is included in the "public" water price of DW (0.6 $\notin m3$), RW (0.2 $\notin m3$) and DRW (0.42 $\notin m3$), but only if it is under 300 m of altitude. As a subsurface irrigation system (SDI), uses the soil as a natural advanced (but not high cost) water treatment, the *in situ* reuse of RW produced by lower tech wastewater plants would provide a valuable resource for the small villages.

A successful study case is presented here. A 2-year SDI experiment was performed using RW, cultivating alfalfa (*Medicago sativa*) and Sudan grass hybrid (*Sorghum bicolor ssp sudanense*). Although a saline (EC: 2.24 dS/m) and sodic (SAR: 6.9) water was used, irrigation with SDI led to high forage yields (between 8.4 and 11.9 kg of fresh matter/m² obtained by 8 harvests per year). However, sand filter and drip lines have to be cleaned weakly and every 6 months respectively. As a conclusion of this study, adequate plant germination and yields, and feasibility of SDI using a secondary effluent was demonstrated. Besides, water cost represents only 24 to 17 % from the gross income.

Key words: advanced treatment, alfalfa, water-cost, reclaimed-water management, SDI, sudangrass hybrids

INTRODUCTION

In the Canary Islands, water scarcity is one of the constraints for agricultural activity. The islands have a volcanic origin with a subtropical climate. The main risk factors for soil degradation identified in Europe are: erosion, organic matter decline, compaction, salinisation and landslides (Eckelmann *et al.*, 2006). Proper land use would prevent these problems but water is necessary to sustain agricultural activity. For instance, the most arid islands (Fuerteventura and Lanzarote) have an annual rain average of 150 $L \cdot m^{-2}$. In Gran Canaria, agriculture uses the 58% on the total consumption, about 75 Mm³ (Consejo Insular de Aguas de Gran Canaria, 1998) and has to compete with other uses, like tourism, which is the basic economical income.

Aquifer overexploitation has led to a decline in water quality: from the total groundwater extracted, the 43 % is pumped from non renewable resources (ground water represents 56% from the total consume while superficial resources or "fresh water" are only 7%; Consejo Insular de Aguas de Gran Canaria, 1998).

The price fluctuation of groundwater and fresh water (0.3 to $0.7 \notin m^3$), which is a private resource in these islands, often limits the recovery of the agricultural investments. As an example, water consumption involves from 30% (tomato) to 45% (banana) of the cost production. Availability of alternative resources at a foreseeable price, will allow people to design the optimal infrastructures adapted to their needs. Alternative resources are reclaimed water, desalinized seawater and desalinized reclaimed water (RW, DW and DRW). Small mountain villages use to transport their effluents to the high-tech wastewater plants near the most populated coastal areas. Due to the high cost of pumping, the RW produced by the treatment plants are located in the lower areas, and then, that water is not available to the people living in the mountains. As a consequence, the pumping cost is included in the "public" water price (without any subsidy) of RW (0.2 m_3) and DRW (0.42 m_3) for lands below 300 m of altitude (Consejo Insular de Aguas de Gran Canaria, 2004), while DW is from 0.5 to 0.6 m_3 (Consejería de Aguas Gobierno de Canarias, 2007). Thus, the *in situ* reuse of RW produced by lower tech wastewater plants would provide a valuable resource for the small villages and would eliminate the need for the costly transport of RW, DW or DRW from the lower areas. Forage production is the most suitable choice to reuse RW resources, due to its lower quality exigency and the lack of the fresh forage crops to feed livestock caused by the high water price (Palacios *et al.*, 2005).

A disinfected secondary effluent could be used for drip irrigation, after a conventional sand media and mesh filtration, whenever a safety reuse is assured by an advanced but not high cost water treatment. Two points of view should be considered in any RW irrigation project: risk and sustainability principles. A subsurface irrigation system, SDI, uses the soil as a "natural advanced water treatment" resulting in a safe and profitable reuse. Design and management are closely linked in a SDI system. Besides, SDI must also have long life in order to be economically viable to produce the relative low value field crops (Rogers and Lamm, 2005), such as forages are. SDI has other advantages: water-conserving production practices, high yielding crops, prevention of the contact between water and stems and leaves minimizing sanitary risks (Camp, 1998). Sodium enrichment in soils affects its structural stability and its permeability (being estimated as hydraulic conductivity) and its holding capacity. Thus, the change in hydraulic conductivity is related to water EC and SAR but it is also affected by many other factors as pH, soil texture and clay minerals, soil organic matter, water organic matter content, and water and soil management. Many studies are published about the estimation on structural stability related to water quality (Suarez et al., 1984, Ayers and Westcot 1985), but according to the numerous factors implied, the threshold levels must be adapted for RW irrigation under SDI management.

While salinity thresholds established for conventional resources are mainly related to Cl and Na, RW carries out many other compounds. Thus, if proper fertilizer application program is calculated, higher EC levels of RW could be applied with no adverse effects on plant production (Palacios *et al.*, 2000). Nitrogen, phosphorous and boron carried out by RW has also a critical effect in soils, crops and environment. Water management due to the irrigation system design and operation affects the medium and long time consequences.

The aim of this work was to evaluate agronomical and economical feasibility of RW reuse for the forage production applied by SDI system and to determinate the best SDI design and water management practices under experimental field conditions as a pilot reuse project.

MATERIAL AND METHODS

A two year irrigation experiment was performed using RW. The experimental field is located on the north coast of the Gran Canaria Island, Spain. The annual precipitation is 243 mm and the average annual temperature is 19.5° C. The effects on the soil and forage (yield and quality) was monitored and compared with a control using conventional freshwater. Thus, the experimental field is divided in two zones, respectively irrigated with reclaimed water (RW) and fresh water (FW). Each water quality zone has its own pond, filtration system (sand and screen mesh) and pipeline network. Alfalfa (*Medicago sativa*) and sudangrass hybrid (*Sorghum bicolor ssp sudanense*Sorghum bicolor ssp bicolor*) were cultivated at the experimental site. No fertilizers were added. A SDI system was used for water distribution. The lateral lines were buried at 0.2 m depth. The experiment consisted

of 8 irrigation treatments: 0.5 and 0.75 m emitter spacing, 0.5 and 1 m between lateral lines, and two types of emitters, both of them pressure compensating. The irrigation frequency depended on the climatic conditions, varying from daily to 3 times per week. Irrigation days, amount of water applied in each plot, and biweekly pressure readings were recorded. Water, plant and soil samples were collected and submitted to the laboratories for analysis. Water from 22 dates (since October 2001 to May 2003) were analyzed. Filtered RW (RWf) was also analyzed for the last 14 dates. Composed soil samples from all plots were collected before seeding and during the experiment, to carry out the chemical analysis and the soil texture determination in the laboratory. Selective soil sampling in June 2003 was performed in order to study salt and nutrient distribution around the emitters: the XY data were obtained from 3 distances apart from the drip line: 0, 0.25 and 0,50 m, and 4 different deeps under the soil surface: 0.1, 0.2, 0.3 and 0.5 m, for the soils irrigated by both RW and FW. Repetitive soil solution samples were obtained during the experiment using "Rhizon" samplers. Plant samples for nutrient determination were weighted directly after harvest on 26/03/03and on 9/05/03 (Sudan grass), and on 28/05/03 (alfalfa). No fertilization program was used for any of the treatments. Forage was harvested by hand, when 10% of the crop was flowering. More detailed aspects of experimental design are presented in Palacios et al., 2005. Microbiological parameters were also determined but presented in another paper.

RESULTS AND DISCUSSION

Water characteristics

Irrigation RW had acceptable BOD₅ values for drip irrigation (16 ± 5 and 12 ± 6 mg·L⁻¹, before and after filtration respectively). However, in September 2002 and from May to August 2003, values greater than 150mg·L⁻¹ were produced due to operation problems in the wastewater treatment plant. In spite of this, effective reduction was obtained after sand filtration since DBO5 values of filtrated RW was continuously below 125 mg L⁻¹. This illustrates the effectiveness of the sand filtration step to protect the irrigation system.

Main water characteristics during the experiment are presented in table 1. As observed, water quality of the two irrigation water treatments were different (P<0.05) in all the measured parameters with the exception of pH (P>0.05). Irrigation RW showed a moderate degree of restriction on use regarding salinity and boron concentration, and no restriction concerning infiltration (Ayers & Westcot, 1985). In other hand, as pointed out by Palacios *et al.* (2000) for bananas, the salinity threshold tolerance for crops due to the RW nutrient contribution is higher than obtained for conventional water resources. Comparing with preliminary results presented in Palacios *et al.* (2005), lower salinity (2.24 from 2.45) and SAR (6.94 from 8.36) are obtained, caused by the improvement of the wastewater plant design and management. A moderate degree of restriction was also shown for boron, nitrate and chloride. In relation to sustainability, the contents of B (0.19 (mmolc·L⁻¹, equivalent to 2.0 mg·L⁻¹), P (0.14 mmol·L⁻¹ or 4.21 mg·L⁻¹), and N (1.66 mmol·L⁻¹ or 23.27 mg·L⁻¹) exceeded the concentration limits proposed for long-term irrigation, respectively, 0.75, 0.05, and 5 mg·L⁻¹ (ANZECC, 2000; USEPA 2004) but are acceptable for short term criteria (2, 0.8-12 and 25-125 mg·L⁻¹ respectively). The rest of the nutrients and metals in RW were acceptable for long-term irrigation.

Soil characteristics and water management control nutrient availability to plants. Factors affecting B availability in soils are solution pH, soil texture (B adsorption depends on the clay content), soil moisture (Perkins in 1995 studied the soil moisture content effects in boron availability), temperature and organic matter content (Goldberg, 1997). Thus, water management will modify plant uptake. In this sense, SDI water management effect on boron availability are not enough studied. Soil, stems and leaves contents presented in this paper will allow us to increase this useful knowledge.

	auring the expe				2	2			
Irrigation	EC	pH	Na^+	\mathbf{K}^+	Ca ²⁺	Mg^{2+}			
Water	dS m^{-1}			$mmolc \cdot L^{-1}$					
FW	0.15 (0.14)	7.15 (1.01)	0.56 (0.91)	0.20 (0.08)	0.60 (0.12)	0.20 (0.19)			
RW	2.24 (0.75)	7.76 (0.56)	14.55 (5.97)	1.40 (0.46)	3.86 (1.24)	4.93 (2.15)			
	TSS	SAR	Cl	SO_4	Alk	NO ₃ ⁻			
	mg·L⁻¹	$(mmol \cdot L^{-1})^{1/2}$		$mmolc \cdot L^{-1}$					
FW	3.03 (3.40)	0.88	0.59(1.12)	0.09(0.12)	1.11(0.98)	0.01(0.01)			
RW	17.87 (3.76)	6.94	12.10(5.45)	3.8(0.76)	9.03(2.24)	0.9(1.3)			
	NH_4	Pt	В	Fe	Mn	Zn			
		mmolc·L ⁻¹			µmol∙L⁻¹				
FW	0.01 (0.01)	DL	DL	0.53 (1.08)	0.14 (0.31)	0.32 (0.31)			
RW	0.77 (0.81)	0.14 (0.09)	0.19 (0.06)	7.84 (23.08)	0.7 (1.25)	3.35 (3.83)			
TSS: suspended solids		Pt: total phosphorus	Alk: alkalinity	DL: Under detection limit					

Table 1. Chemical composition of RW and FW. Mean and (standard deviation) values from 21 waters samplings during the experiment.

Soil

The field plot was an old tomato orchard, in fallow for more than 20 years. The soil was a clay loam (27.6% clay), originally saline and sodic (ESP 24%) and could be classified as Anthrosol (FAO, 1998) or Torriarent (Soil Survey Staff, 1998)..

Soil characterization during the experiment is presented in table 2. As observed, very high values of EC in soils irrigated by RW were obtained in the summer of 2003. This soil salinity is above the threshold limit even for tolerant crops as alfalfa and sudan grass hybrid. As expected, lower salinity and nitrate levels were found in soils irrigated by FW. As alfalfa is a legume and the sudan grass hybrid was not cultivated in FW irrigated soil, no N fertilization was needed. Organic matter content was higher in FW soils as compared to RW. This apparently contradictory result is consistent with other authors (Magesan et al., 1999). The soil can be classified as rich regarding phosphorus contents. Thus, the crops were able to absorb enough P without using any P fertilizer, as leaves contents were in the acceptable interval. However, the greater P Olsen contents measured in FW soils should be further studied.

Table 2: Evolution of soil properties during the experiment

	II OI	bon pro	perties au	ing the e	Aperiment							
Date	09/2001		04/2002	11/2002	02/2003	06/2003		07/2003		12/2003		
Irrigation Water						RW	FW	RW	FW	RW		
EC _{s*}			5.02	4.26	3.81			15.33	2.69	8.25		
pH-water			7.9	7.9	8.3	8.0	7.9	7.8	8.2	7.8		
Ca Carbonate (%)			4.6	5.5	4.0	3.3	4.0	4.4	4.8	4.2		
Organic C (%)			0.56	0.52	0.19	0.41	0.8	0.51	0.7	0.62		
C-D Fe**			0.96	0.89	0.33	0.71	1.36	0.88	1.26	1.07		
Ex Cations***	Κ	4.6	4.0	4.7	3.8	5.0	7.4	6.5	5.5	4.3		
	Na	7.0	10.1	8.3	7.6	9.5	18.5	23.6	8.8	8.8		
	Ca	8.0	12	12.8	12.8	14.8	16.9	15.5	14.9	12.8		
	Mg	8.8	7.9	9.0	8.5	9.1	10.2	10.7	9.7	9.5		
Olsen P(mg/kg)		28	41	40	22	27	64	32	48	36		
NO ₃ -N _{S****}				250	97	90		120	25	200		
*EC _s :electrical	Saturated Extrac				**C-	**C-D Fe: Citrate-			****Ex Cations (cmol _c .kg ⁻¹)			

conductivity

Dithionitre extractable Fe

 $****NO_{3}-N_{S}(mg/kg)$

The selective soil sampling EC results measured at the end of the experiment were drawn by SURFER 8 (Golden Software, 2002) software from the XY data for the soils irrigated by both RW and FW. Thus, the figure 1 shows the Electrical Conductivity (EC) distribution in the soil profile. Different scale was used due to the high differences of EC values between RW and FW soil, caused by the salinity of the RW.

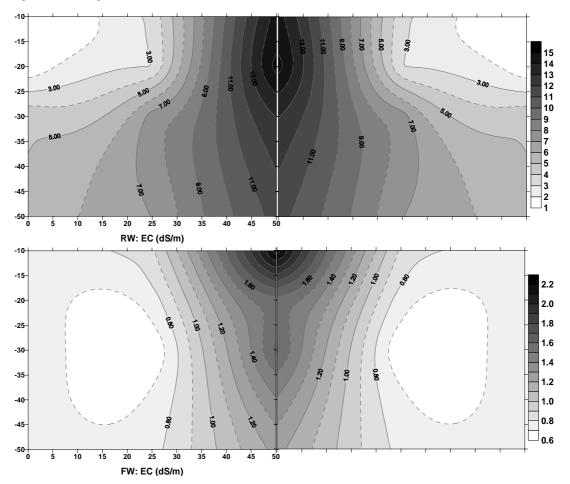
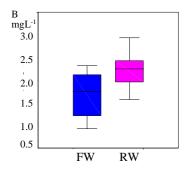


Figure 1: Electrical Conductivity (EC) in the soil profile. XY data were obtained from 3 distances apart from the drip line: 0, 0.25 and 0,50 m, and 4 different deeps under the soil surface: 0.1, 0.2, 0.3 and 0.5 m). Symmetric figure is also drawn to represent the whole profile, as the drip lines are buried 1 m apart from each other.

As expected, higher salinity is obtained in RW irrigated soils. Salinity distribution coincides with results obtained by Assadian *et al.*, (2005) who mentioned higher EC values in soils subirrigated using RW at 0.1 m below the soil surface. As observed, water management related to the SDI system governed the salinity profile: higher EC values were measured at the top and between the drip lines. EC values between lines coincided with those obtained by conventional soil sampling. This fact was also confirmed by the negative effect of salinity on the crop yields.



Soluble boron concentration was higher in RW soils than in FW, as presented in figure 2. In spite of this, no adverse effect was observed in crops therefore alfalfa and sudan grass hybrid are boron tolerant species.

Figure 2: Average Boron concentration in soil solution obtained from Rhizon samplers in RW and FW plots.

Crops irrigated using a SDI system

Previous studies have demonstrated the feasibility of alfalfa production irrigated by SDI system (Hutmacher *et al.*, 1992; Mead *et al.*, 1992; Mc Gill y Hutmacher, 1993 and Hutmacher *et al.*, 1996). A decrease in irrigation water use under SDI was reported when this was compared with spray irrigation on grain sorghum (Colaizzi *et al.*, 2003), using 50 and 25% levels of water applied.

As previously mentioned in Palacios *et al.*, (2005) the water salinity did not affect first germination and shooting that were adequate in all irrigation treatments of our study. It is contradictory with Lamm (2002) who mentioned that salinity could be a problem under this irrigation system and reduce the alfalfa germination rate. That result was obtained in spite of no additional irrigation was used, although it is mentioned to be necessary (Camp 1998). It is possible that the unusual heavy rainfalls had some role in mitigating high salinity of the soil. A heavy weed growth (even when a chemical control was applied) required reseeding of the alfalfa plots from the second and the third blocks. Seeding again was possible since the irrigation system was under the ground and allowed to sow without inducing any disturb. That should be considered as another advantage of this SDI system.

Figure 3 shows the yield for alfalfa and sudan grass hybrid during the year 2003. The first alfalfa harvest had to be rejected due to weed infection. Water quality was significant for alfalfa production except for the most productive month (June) for RW and the lowest one (November) for FW. Thus, the yields are represented separately for FW and RW. The greater difference by water quality was found in may 2003, coinciding with the highest salinity obtained in RW soils. Although alfalfa is a salinity tolerant crop, the high EC values obtained between the irrigation lines had affected alfalfa production.

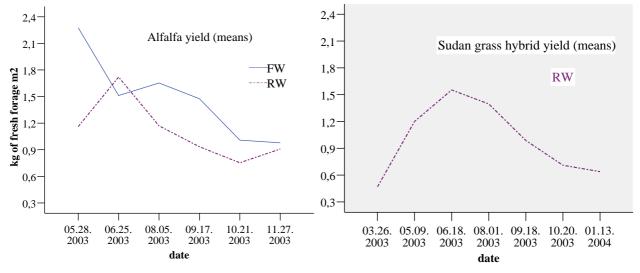


Figure 3: yield for alfalfa and sudan grass hybrid during the year 2003

As expected, statistical differences were obtained by date for both crops, producing the highest yield in spring and the lowest in winter. Similar production was obtained for both crops in summer (not very hot in Canary Islands) but slightly lower production was obtained by sudan grass in winter (alfalfa has a C3 metabolism and sudan grass hybrid is a C4). In fact, sudan grass had no production from 20 October to January while alfalfa had another harvest in November 27, although with the lowest yield. Our mean value per harvest for sudan grass irrigated using RW was 1.05 kg of fresh forage per m². Considering 8 harvest per year (no cut from December to February), the accumulated yield for a year is 8.39 kg·m² (17 Mg of dry matter per ha⁻¹). Higher sudan grass hybrid yield was mentioned by Serra *et al.*, (2005) who obtained a mean value for two years of 25.5Mg·ha⁻¹ of dry

matter, although they harvested only 2 times by each year in the north of Spain. These authors do not mention the quantity of irrigation water used for their study and had small plots of 6 m^2 that probably affected the results.

The soil moisture distribution obtained for all the irrigation treatments was sufficient for effective plant growth. In our study water moved horizontally more than 0.6 m from the drip lines and reached the soil surface above the lateral line. In fact, the significantly higher yield observed in the first cutting in the closer lateral spacing (0.5 m as compared to 1 m) disappeared after the second alfalfa cut. These results for alfalfa production are consistent with obtained by Alam *et al.*, (2002) in semi-arid Kansas, who concluded that it is more economical to use the 1 m spacing. The same results were obtained for sudan grass. No significant differences were obtained from the differences by the emitter spacing (0.5 and 0.75 m), nor by the two types of emitters (two commercial irrigation materials adapted for irrigation using RW).

The SDI system maintenance was monitored: sand filter and drip lines have to be cleaned weakly and every 6 months respectively when irrigating with RW. Thus, effective filtration was able to protect the irrigation system from clogging was obtained using conventional equipment.

As a conclusion, water quality, date and its interaction give an explanation model for the obtained alfalfa yield with a R^2 of 0.65. All of these variables are significant for the model.

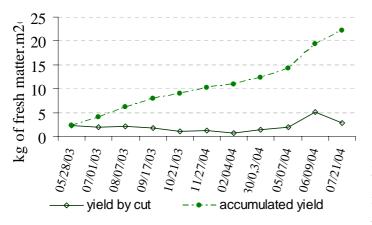


Figure 4: alfalfa production by harvest and accumulated yield irrigated using FW during 14 months

Figure 4 represents the alfalfa yields (in kg of fresh matter per m²) obtained for each harvest and accumulated yield during a more extended period of time (14 months). As presented here, it is possible to obtain 9 harvests per year, and to reach an annual yield of 143 Mg of fresh matter per hectare (equivalent to 28.6 Mg of dry matter·ha⁻¹) irrigating with non limiting water quality. Water consumption was measured by the installed flowmeters for each treatment. All of them used 600 L·m⁻² per year which

Although a saline (EC: 2.24 dS/m) and sodic (SAR: 6.94) water was used, irrigation with SDI resulted in an effective production of forage (between 8.4 and 11.9 kg of fresh matter/m² obtained by 8 or 9 harvests per year). As demonstrated, using FW the yields can be even higher. The salinity problems found in RW irrigated soils negatively affected the crop production. Lower salinity water should be used if the maximum yields are needed. However, enough production is obtained using the RW quality presented in this case study. Neither nutrient deficiencies nor imbalances were observed, suggesting the contribution from irrigation water and probably from residual soil forms.

The price of hey forage is $0.24 \notin kg$. Considering the lower estimated annual yield of 8.4 kg of fresh matter·m⁻² (equivalent to 2.1 dry matter.m⁻²), the gross income is $0.5 \notin m^{-2}$. As RW cost is $0.2 \notin m^{-3}$ and 600 l·m⁻² are applied by SDI system, it can be calculated that the cost of RW represents less than 24 % of the total income. Thus it is possible to cultivate many abandoned soils in Canary Island to obtain enough forage for animal production. Obviously, obtaining a higher forage yield

(11.9 kg of fresh matter \cdot m⁻²) than that presented for alfalfa or when harvesting nine times per year, the ratio will be even lower (17%).

CONCLUSIONS

The proposed water, soil and crop production system is able to produce the forage even in the winter in Canary Islands. Due to the high water volumes required for the fodder crop production, only the use of RW presents a possibility for sustaining this activity in the island. In this study adequate crop emergency, performance and yield, and feasibility of SDI when reusing reclaimed water was demonstrated as no adverse effect is observed in the irrigation system. The farming land operations are compatible with the installation of the subsurface drip irrigation. The cost of RW represents less than 24 % of the gross income even using this salinity water. Higher yield should be expected reducing the RW salinity

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